1 Response Surface Model Based Emission Source Contribution and

2 Meteorological Pattern Analysis in Ozone Polluted Days

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18 ABSTRACT

Urban and regional ozone (O_3) pollution is a public health concern and causes damage to 19 ecosystems. Due to the diverse emission sources of O₃ precursor and the complex interactions 20 of air dispersion and chemistry, identifying the contributing sources for O₃ pollution requires 21 integrated analysis for guiding emission reduction plans. In this study, the meteorological 22 characteristics leading to O₃ polluted days (in which the O₃ levels are higher than the defined 23 standard) in Guangzhou (GZ, China) were analyzed using the 2019 data. The O₃ formation 24 regimes, source apportionments under various prevailing wind directions were evaluated using 25 a Response Surface Modeling (RSM) approach. The results show that O₃ polluted days in GZ 26 2019 can be classified into four types of synoptic patterns (i.e., cyclone, anticyclone, trough, 27 and high pressure approaching to sea) and are strongly correlated with high ambient 28 temperature, low relative humidity, week wind speed, variable wind directions. Additionally, 29 the cyclone pattern strongly promotes O_3 formation due to its peripheral subsidence. The O_3 30 formation is NO_x-limited under the northerly wind while VOC-limited under other prevailing 31 wind directions in GZ. Anthropogenic emissions contribute largely to the O₃ formation of GZ 32 (54-78%) under the westerly, southwesterly, easterly, southeasterly, or southerly wind, but only 33 moderately (35-47%) under the northerly or northeasterly wind. Furthermore, local emissions 34 have the largest O₃ contributions (39-60%) regardless of prevailing wind directions, especially 35 the local NO_x contributions (19-43%). The dominant upwind regional emissions contributed to 36 12-46% of the O₃ formation in GZ (e.g., DG contributions are 12-20% under the southeasterly 37 wind). The emission control strategies for O₃ polluted days should focus on local emission 38 sources in conjunction with the emission reduction of upwind regional sources. 39



42 **1 Introduction**

Over the past five years, the particulate matter (PM2.5) levels have been reduced by 34% 43 in the city cluster of the Pearl River Delta Region (PRD) in China, while the ambient ozone 44 (O_3) concentrations have shown an upward trend. This air quality transition points O_3 as the 45 primary pollutant in the region from 41% in 2015 to 67% in 2020 (DEEGP, 2021). In China, 46 O₃ polluted days are those that have the maximum daily 8-h average (MDA8) O₃ concentration 47 higher than the China Class II National O₃ Standard (160µg/m³) and will cause serious damage 48 to ecosystems. To lower the MDA8 O3 concentrations, enhanced emission control during O3 49 polluted days has been consistently requested by decision-makers in recent years. However, 50 due to the complexity of O₃ formation in ambient air and the diverse source contribution of O₃ 51 precursors, there is still a knowledge gap to formulate emission control policies for O₃ control 52 in the region (Stevenson et al., 2013; Wang et al., 2009). 53

Earlier studies have apportioned the contributing emission sources leading to ambient O₃ 54 pollution using the tagged tracer techniques in Air Quality Models (AQMs) (Ge et al., 2021; Li 55 et al., 2013; Li et al., 2012; Shu et al., 2020). These works mostly focused on specific O₃ seasons 56 and therefore did not adequately reveal the source contribution characteristics in relatively short 57 periods of O₃ pollution. Besides, the tagged tracer techniques have limitations in representing 58 the nonlinear relationship among O_3 and its precursors, especially NO_x (Chatani et al., 2020). 59 Our recent developments of Response Surface Models (RSMs) has been demonstrated for their 60 capability to quickly estimate the nonlinear response of O_3 to its precursors for emission source 61 contribution analysis (Fang et al., 2020; Kelly et al., 2021; Xing et al., 2018; Xing et al., 2020; 62 You et al., 2017). For instance, Xing et al. (2020) employed the RSM technique to identify the 63 O_3 formation regime and indicated that NO_x emission reductions will increase the O_3 64 concentrations in the North China Plain under VOC-limited conditions. Fang et al. (2020) used 65 the polynomial functions-based extended RSM technique integrated with differential methods 66 3

(pf-ERSM-DM) to analyze the dynamic O₃ source contributions over the PRD and projected that the enhanced NO_x control can effectively lower the ambient O₃. However, these studies were also conducted in specific O₃ seasons which did not sufficiently incorporate analyses on the characteristics of meteorological patterns for elucidating the effects of prevailing wind directions that influence O₃ source apportionment results and account for the effect of upwind regional emissions and long-range transport (Berezina et al., 2020; Shao et al., 2009; Suthawaree et al., 2008).

Accordingly, taking Guangzhou (GZ, a representative megacity in the PRD) as an example, this study aims to advance the understanding of regional O_3 pollution through analyzing the meteorological characteristics leading to O_3 polluted days in GZ based on the 2019 data, then using the advanced Response Surface Modeling approach to identify the O_3 formation regimes, source apportionments under various prevailing wind directions. The analysis method and results are expected to assist decision-makers to formulate sound emission control strategies for O_3 polluted days and improve the O_3 attainment rate.

81 2 Methodology

The process of this study is shown in Fig. 1. Firstly, the O₃ polluted days of GZ 2019 are 82 selected based on the monitoring data and the corresponding historical meteorological data (i.e., 83 weather chart, temperature, relative humidity, wind speed, and prevailing wind direction) are 84 also collected. Then the synoptic patterns in O₃ polluted days are classified by the weather 85 charts and the corresponding meteorological characteristics under each synoptic pattern are 86 analyzed. Secondly, the O₃ polluted days are divided into several O₃ episodes with various 87 prevailing wind directions, and the control matrix for these O₃ episodes is designed based on 88 the selected control factors (i.e., pollutants, regions) and parameterized by the 2019 PRD 89 emission inventory (EI) to obtain scenario runs (Xing et al., 2011). Thirdly, the Community 90 Multi-scale Air Quality model (CMAQ) simulations under various scenarios are driven by the 91 2019 PRD meteorology in O₃ episodes derived from the Weather Research and Forecast model 92 (WRF), thereafter the simulation results are utilized to build the pf-ERSM-DM system. Finally, 93 the O₃ formation regime is analyzed and the O₃ source apportionment is evaluated for 94 identifying the major emission source contributors of O₃ under multi wind directions, regions, 95 and precursors based on the established pf-ERSM-DM system (Fang et al., 2020). 96

97 **2.1 Observation data source**

The MDA8 O₃ concentrations of GZ are calculated by the average of that in the ten local 98 national-controlled air-monitoring sites, and the locations of these sites are shown in Fig S1. In 99 2019, GZ suffered apparent O_3 pollution with the yearly 90th percentile of the MDA8 O_3 100 concentration reaching the peak over the past five years (Fig. S2), and a total of 54 O₃ polluted 101 days happened as listed in Table S1 (the monitoring data are accessed in Guangdong 102 Environmental Quality platform: http://113.108.142.147:20061/StationStatus/AppCheck). 103 Accordingly, for analyzing meteorological patterns and emission source contributions in O₃ 104 polluted days of GZ, the 2019 data was selected in this study. 105

The weather charts which are utilized to classify the synoptic patterns of O₃ polluted days 106 GΖ 2019 obtained from the Hong Kong 107 in are Observatory (http://www.weather.gov.hk/tc/wxinfo/currwx/wxcht.htm). The main synoptic patterns of 108 109 those polluted days can be divided into four types, namely cyclone, anticyclone, trough, and high pressure approaching to sea according to the classifications of the former researches 110 (Huang et al., 2005; Pan et al., 2017). The typical examples of weather charts on each synoptic 111 pattern are shown in Fig. S3. Meanwhile, the meteorological factors of O₃ polluted days in GZ 112 2019 including temperature, relative humidity, wind speed, and prevailing wind direction that 113 can significantly affect the ambient O₃ (Han et al., 2020; Shao et al., 2009; Shi et al., 2020; 114 Wang et al., 2017) are also collected. These meteorological data are acquired from the National 115 Climate Data Center (https://www.ncei.noaa.gov/). 116

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2.2 WRF-CMAQ simulation system

The O₃ episodes with various prevailing wind directions are listed in Table S2 which are 118 selected based on the polluted days. The WRF version 3.9.1 is applied for providing the 2019 119 PRD meteorological inputs in the O₃ episodes for the CMAQ model and the CMAQ version 120 5.2 is utilized to simulate the O_3 concentration. Three nested domains (i.e., d01, d02, d03) used 121 for the WRF-CMAQ model are shown in Fig. 2 and the horizontal resolutions of the three 122 domains are 27km, 9km, 3km with the grid cells of 175×124 , 133×133 , 127×160 respectively. 123 The innermost domain (d03) covering the entire PRD is divided into 8 regions, namely 124 Guangzhou (GZ), Foshan (FS), Zhongshan (ZS), Jiangmen (JM), Dongguan (DG), Shenzhen 125 (SZ), Huizhou (HZ) and the other regions in the d03 domain (OTH). For the d01 and d02 126 domains, the emission inventories offered by Tsinghua University are adopted (Multi-127 128 resolution Emission Inventory For China, MEIC: http://meicmodel.org). The 2019 PRD emission inventory of the d03 domain (Table S3) is developed by our research group. The 129 Model of Emissions of Gases and Aerosols from Nature (Megan) version 2.10 (Guenther et al., 130

2012) is utilized to prepare the natural emissions. The initial and boundary files of the d02 and
d03 domain are from the CMAQ simulation results on the upper domain and a 5-day spin-up
time is used to rid of the initial condition influence. The detailed parameter configurations for
WRF and CMAQ are shown in Table S4 and Table S5. The performance of WRF and CMAQ
is detailed in section 3 of Supporting Information (SI).

136 **2.3 Response surface modeling technology**

RSM is a meta-model established from the CMAQ simulation results with the ability to 137 quickly and accurately estimate the pollutant response to emission source perturbations. The 138 pf-ERSM-DM system is used in this study, which has been elaborated on in our latest paper 139 (Fang et al., 2020; 2021). The experiment design for constructing pf-ERSM-DM is presented 140 in Table S6 and the matching control matrix is the same for each O₃ episode as shown in Fig. 141 S4. Summarily, there are 181 scenario runs of CMAO simulations sampled by the Hamersley 142 quasi-random Sequence Sampling (HSS) with the emission ratio ranges from 0.0 to 1.5, 143 including one baseline scenario (=1.0), 8 single regional RSMs (for the d03 domain being 144 divided into 8 regions), and one multiple regional RSM. For a single regional RSM, 20 145 scenarios are established for the response of O₃ in the individual receptor region to precursor 146 emission changes in the individual source region, and emissions from the other source regions 147 are consistent with the baseline. For multiple regional RSM, 20 scenarios are created for O₃ 148 response in individual receptor regions to simultaneous precursor emission changes from all 149 source regions. Additionally, 10 extra scenarios which are also sampled by HHS are used for 150 out-of-sample (OOS) validation to evaluate the accuracy of the established pf-ERSM-DM 151 system, and the OOS validation results are detailed in section 3 of SI. 152

153 **3 Results and discussion**

154 **3.1 Meteorological characteristics under various synoptic patterns in O₃ polluted days**

To analyze the meteorological patterns associated with O_3 polluted days in GZ 2019, four main synoptic patterns and their corresponding meteorological characteristics are summarized in Table 1.

The cyclone pattern has the highest occurrence frequency of 26 days, mainly happening 158 in summer and autumn. Under this pattern, it exhibits a high temperature (mean value is 31°C) 159 and a dry air layer (mean relative humidity is 52%) in the low troposphere which are affected 160 by the thermal effects-driven from the strong subsidence airflow at the outskirts of this pattern, 161 causing the average value of MDA8 O₃ concentration (200µg/m³, with the range of 165-162 248µg/m³) significantly higher than that under the other three synoptic patterns. The 163 anticyclone pattern occurred in spring, summer, and autumn, with the frequency of 14 days, in 164 which the accumulation of pollutants is promoted at the periphery of this pattern. Compared 165 with the other three patterns, the mean temperature (27°C, with the range of 17-34°C) is 166 relatively low and the mean wind speed (1.32m/s) is at a comparatively high level which may 167 cause the average value of MDA8 O_3 concentration (187µg/m³, with a range of 163-228µg/m³) 168 of this pattern is the lowest. The trough is situated in the vicinity of the east over the South 169 China Sea happening for 9 days in autumn, under which the average value of MDA8 O₃ 170 concentration (190µg/m³) is ranked third among the four synoptic patterns. The mean wind 171 speed is relatively weak (1.25m/s), while the mean relative humidity (54%, with the range of 172 30-80%) is higher than that under the other synoptic patterns. The high pressure approaching 173 to sea pattern occurred the least days of 5 in spring and autumn. Under this circumstance, GZ 174 is located at the bottom of the high-pressure ridge, in which the mean wind speed (1.08m/s, 175

with a range of 0.3-2.4m/s) is the lowest among the four synoptic patterns, and the average value of MDA8 O_3 concentration (192µg/m³) under this pattern is the second-highest.

Moreover, it is noticed that the prevailing wind directions under all synoptic patterns are generally variable. Due to the fact that GZ is located in the center of the PRD surrounded by abundant emission sources, the wind blowing from any direction may bring the transmission of pollutants to the local and influence its O_3 formation. Hence, further analysis of emission source contribution in O_3 polluted days of GZ under various prevailing wind directions is needed and detailed in the following sections.

184 **3.2 Analysis of O₃ formation regime**

The 2-D isopleth of O_3 is plotted with the emission ratios of VOC and NO_x as the X and 185 Y axes, and the combination emission ratios of VOC and NO_x that produce the same O₃ 186 187 concentrations are connected through isolines (Li et al., 2022). This isopleth can provide useful information about the O₃ formation regime (i.e., VOC-limited or NO_x-limited regime) which 188 reflects the nonlinearity between O₃ and its precursor emissions (Fang et al., 2020; Jin et al., 189 2021). In our previous research, the peak ratio (PR), which is described in Text S4, has been 190 used as the indicator to determine the O₃ formation regime (Xing et al., 2018). In general, the 191 PR value lower than 1 suggest that the O₃ formation is VOC-limited, otherwise, it is NO_x-192 limited. To analyze the O₃ formation regime in O₃ episodes of GZ 2019, the 2-D isopleths of 193 O₃ response (The average value of MDA8 O₃ concentrations in O₃ episodes of GZ under various 194 prevailing wind directions) in GZ to Anthropogenic VOC (AVOC) and NO_x emission changes 195 from all source regions in the d03 domain are obtained and the corresponding PR values are 196 also calculated, as shown in Fig. 3. 197

When the same wind prevails, the trend of 2-D isopleths is analogous and the PR values exhibit little difference with the range of 0-0.14. However, the PR value of O_3 episode in summer is higher than that in other seasons under the same prevailing wind, which is due to the

201 seasonal variation of Biogenic VOC (BVOC) emissions while NO_x emissions vary slightly in different seasons (Fang et al., 2020; Shao et al., 2009; Yang et al., 2019). It can also be found 202 that the average PR values under the northerly, westerly, northeasterly, southwesterly, easterly, 203 southerly, southeasterly wind are 1.08, 0.80, 0.79, 0.69, 0.60, 0.48, 0.44 respectively. The 204 average PR value is the highest under the northerly wind since the BVOC emissions are 205 relatively plentiful in the north of GZ while the NO_x emissions are comparatively small as 206 shown in Fig. S5, which may facilitate the O₃ formation in GZ to the NO_x-limited regime 207 because of the dominant reaction of NO with peroxyl radicals. When the southeasterly wind 208 prevails, the NO_x emissions in local and upwind cities are abundant, while the VOC emissions 209 are relatively small. In this case, the main removal process of free radicals is their reaction with 210 NO₂ to forms the nitric acid, hence the O₃ formation will be mainly determined by the 211 212 generation rate of free radicals, resulting in the strong VOC-limited regime and the lowest average PR value. 213

Under the northerly wind, the O_3 formation in GZ is NO_x -limited (1.03 < PR < 1.11), 214 indicating that the priority control of NO_x (even at slight reductions) will have a more obvious 215 effect on O₃ mitigations while the influence of VOC reduction is comparably tiny. Under the 216 other prevailing wind directions, the O_3 formation in GZ is generally VOC-limited (0.38 < PR 217 < 0.82), implying that an initial VOC control is conducive to reducing the baseline O₃ levels in 218 GZ, whereas the NO_x emission reductions will increase the O₃ concentration initially because 219 of the competition between NO₂ and VOC for OH radicals. Only the NO_x emission reduction 220 rate greater than 18-62% can make the O₃ formation regime be transferred to the NO_x-limited 221 condition and lower the O₃ concentrations in GZ. Additionally, to prevent the phenomenon of 222 223 NO_x emissions being reduced but the O_3 concentration is rising, the minimum VOC/NO_x reduction ratios (VNrs) are also calculated and described in Text S4 (Xing et al., 2018). 224 Accordingly, the VNrs are 1.13-1.15, 1.20, 1.33, 1.33-1.53, 1.43-1.47, 1.45-1.87 under the 225 10

westerly, northeasterly, southwesterly, easterly, southerly, southeasterly wind, respectively. The results show that the VNr is negatively correlated with the PR which indicates that a stronger VOC-limited regime requires more simultaneous VOC control with NO_x; besides, the VNrs fluctuate with the change of the prevailing wind directions, which needs to be considered in emission control policy formulation.

3.3 Analysis of O₃ source apportionments

232 **3.3.1** Anthropogenic and background contributions to O₃

Differentiating of the anthropogenic contribution (anthropogenic emission sources of 233 VOC and NO_x within the d03 domain) and background contribution (including initial condition, 234 boundary condition, and the non-human controllable emission sources of VOC and NO_x) (You 235 et al., 2017) to O₃ can offer information about the maximum achievable O₃ concentration targets 236 and support for formulating effective O₃ control strategies. Through the established pf-ERSM-237 DM system, the anthropogenic and background contributions to O₃ in GZ (The average value 238 of MDA8 O₃ concentrations in O₃ episodes of GZ under various prevailing wind directions) 239 can be quantitatively analyzed when eliminating all anthropogenic emissions as shown in Fig. 240 4. 241

Under either the westerly or the easterly wind, the occurrences of O₃ polluted days are few 242 with 5 days. The base O₃ concentrations under the westerly wind are relatively high with a 243 range of 202-214 μ g/m³ while that under the easterly wind (162-176 μ g/m³) are comparatively 244 low. Under these two prevailing wind directions, the anthropogenic emissions have large O₃ 245 contributions accounting for 65-78%, and the base O₃ concentrations can be reduced to the 246 greatest extent of 44-67 μ g/m³. When the southeasterly wind prevails, the O₃ episodes occurred 247 248 the most frequently and the base O_3 concentrations vary in a wide range of $165-219\mu g/m^3$. In this condition, the anthropogenic contributions account for 54-63%, which are lower than that 249 under the westerly or easterly wind, and the base O₃ concentrations can be reduced to 69-250

90µg/m³. These relatively low anthropogenic contributions under the southeasterly wind may 251 be explained by the ocean inflow transported upstream. Similar to the situation under the 252 southeasterly wind, the anthropogenic contributions under the southwesterly or southerly wind 253 are 55-63%, and thus the base O_3 concentrations can be reduced from 170-203µg/m³ to the 254 background O_3 concentrations of 66-84µg/m³. The northeast and north wind usually prevail in 255 autumn, in which the base O_3 concentrations are also high with the range of 172-206µg/m³, 256 while the anthropogenic contributions are less than 50% overall. Under these two conditions, 257 the O₃ concentrations of GZ are significantly influenced by the emissions transported from the 258 continental regions of China, resulting in high background O_3 concentrations (104-120µg/m³) 259 which are above China Class I National O₃ Standard (100µg/m³). Therefore, the implementation 260 of stricter control policies on a larger spatial scale rather than only the PRD may be needed to 261 262 further decrease the O₃ concentrations in GZ under the northeasterly or northerly wind.

Additionally, according to the previous O_3 source contribution analysis conducted in entire months fitted by the pf-ERSM-DM system, the anthropogenic contributions to the O_3 in GZ are 32% and 45% under northeasterly and southwesterly wind respectively (Fang et al., 2020). While, the anthropogenic contributions in O_3 episodes estimated in this study increase to 47% and 63% when the northeasterly and southwesterly wind prevail separately, the anthropogenic contributions matter a lot than that in entire months may be due to the relatively stagnant weather conditions in O_3 episodes.

3.3.2 Anthropogenic contributions to O₃ under multi wind directions, regions, and precursors

To determine the anthropogenic contributions to O_3 under multi wind directions, regions, and precursors, the total NO_x and VOC contributions from multiregional emission sources to the O_3 in GZ are calculated (Table S10). The NO_x and VOC emission contributions from each source region in the d03 domain to the O_3 in GZ under various prevailing wind directions are furtherly estimated, as shown in Fig. 5 and detailed in Table S11.

Considering the total contributions of multiregional NO_x and VOC emissions, it can be 277 seen that NO_x makes more significant contributions (52-80%) than VOC under all prevailing 278 wind directions except under the southeasterly wind in which the total NO_x contributions (46-279 48%) are slightly lower than the VOC in some O_3 episodes. Additionally, the total NO_x 280 contributions under various prevailing wind directions show a declining trend as follows: north 281 (73-80%), west (68-70%), northeast (66%), southwest (63%), east (55-58%), south (52-56%), 282 southeast (46-55%), and this trend may be elucidated by the variation of the corresponding O_3 283 formation regimes. For example, the O_3 formation in GZ is typical NO_x-limited (1.03 < PR < 284 1.11) under the northerly wind when the depth NO_x emission reductions can significantly lower 285 286 the O_3 concentrations, resulting in the largest total NO_x contributions.

GZ is an urban and prosperous city located in the central part of the PRD, with abundant 287 O_3 precursor emissions, as for the contributions from each source region to the O_3 in GZ, the 288 local contributions (39-60%) are always the highest regardless of prevailing wind directions. 289 Furthermore, compared with the local contributions of an entire month (12-17%) in the former 290 study evaluated by the pf-ERSM-DM system (Fang et al., 2020), the local contributions in O₃ 291 episodes obviously increase more than 20% and these similar results have also been proved in 292 other related researches (Li et al., 2012; Yang et al., 2019). This phenomenon may be explained 293 by the conductive meteorological conditions in O₃ episodes which favor the chemical formation 294 of O₃ and lead to considerably high local contributions. Additionally, the GZ-NO_x contributions 295 (19-43%) are higher than GZ-VOC contributions (11-28%) in most O₃ episodes, this is because 296 the O₃ formation in GZ will transit to NO_x-limited when strengthening the emission reductions, 297 hence the influence of enhanced NO_x reduction on O_3 will be more obvious than that of VOC. 298

299	Apart from the local contributions, the major regional emission source contributors in
300	different O ₃ episodes are closely related to the prevailing wind direction. Under the northerly
301	wind, the O ₃ and its precursors may be transported from the northern PRD (i.g., Qingyuan) to
302	GZ, thus, $OTH-NO_x$ emissions show an obvious impact of 26-37% and OTH-VOC
303	contributions are also great accounting for 7-12%. Similar to that under the northerly wind,
304	OTH-NO _x contributions account for a large amount of 19% under the northeasterly wind. When
305	the westerly wind prevails, FS-NO _x and FS-VOC contribute a substantial percentage of $13-17\%$
306	and 7-10% respectively to the O ₃ in GZ, because FS is an essential industrial/manufacture base
307	and a comprehensive transportation hub in the PRD with intensive VOC and NO _x emissions;
308	besides, OTH-NO _x contributions also account for a relatively large proportion of 10-15%.
309	Under the southwesterly wind, FS-NO _x has a large O_3 contribution of 13%; OTH-NO _x also
310	occupies a nonnegligible proportion of 11% which may mostly be from Zhaoqing (an important
311	transportation hub that connects economically developed regions with southwestern provinces)
312	and Zhuhai (a major coastal port of China). Under the southerly wind, FS also has a great
313	impact, with FS-NO _x and FS-VOC contributions accounting for 10-13% and 7-8% respectively.
314	DG is a city with relatively rapid urbanization in the PRD, containing large amounts of
315	industrial factories and vehicles, thus under the southeasterly wind, $DG-NO_x$ and $DG-VOC$
316	contributions are great accounting for 5-11% and 6-11% respectively; OTH-NO _x also has an
317	important impact of 5-10%, which may result from the intensive emissions in the coastal areas
318	of OTH. Resemble the situations under the southeasterly wind, when the easterly wind prevails,
319	DG-VOC and DG-NO _x contributions account for 8% and 6-7%, respectively; the contributions
320	from OTH increase slightly, with OTH-NO _x contributions accounting for 7-8%.

321 **4** Conclusions

In this study, the meteorological characteristics leading to O_3 polluted days in GZ were analyzed in a typical year (2019) with frequent O_3 episodes, and the O_3 formation regimes and source apportionments in O_3 episodes under various prevailing wind directions are comprehensively evaluated based on the pf-ERSM-DM system.

The ambient O_3 pollution is strongly related to meteorological patterns. The O_3 polluted days in GZ 2019 can be classified into four major synoptic patterns (i.e., cyclone, anticyclone, trough, and high pressure approaching to sea) and they generally exhibit similar meteorological characteristics favoring the chemical production of O_3 in GZ, including high mean temperature (26-31°C), low mean relative humidity (48-54%), weak mean wind speed (1.08-1.34m/s) and variable prevailing wind directions. Additionally, the cyclone pattern strongly promotes O_3 formation due to its peripheral subsidence.

From the perspective of emission control in O₃ polluted days, it is momentous to identify 333 the O₃ formation regime and quantitatively comprehend the precursor emissions in shaping 334 anthropogenic and background contributions. The O₃ formation regime and source 335 apportionment results are highly related to the prevailing wind directions due to the abundant 336 pollutants transported from the anthropogenic emission sources in the upwind areas. Under the 337 westerly, southwesterly, easterly, southeasterly, or southerly wind, the O₃ formation in GZ is 338 339 generally VOC-limited and the anthropogenic emissions have large O₃ contributions (54-78%); while under the northeasterly or northerly wind, the O₃ formation is weak VOC-limited and 340 NO_x-limited respectively, and the anthropogenic contributions are relatively moderate (35-47%) 341 resulting in comparatively high background O₃ concentrations which even exceed the China 342 Class I National O₃ Standard (100µg/m³). In terms of anthropogenic contributions, the local 343 contributions account for the largest proportion (39-60%) regardless of prevailing wind 344

directions, especially the GZ-NO_x which occupies a high contribution ratio (19-43%); upwind regional emission sources also had large O_3 contributions (12-46%).

Consequently, the sound emission control strategies for O₃ polluted days should be 347 348 implemented with the consideration of prevailing wind direction. First, the NO_x control (even at slight reductions) should be strengthened under the northerly wind; while the short-term 349 coordinated control of VOC and NO_x and long-term enhanced control of NO_x are suggested 350 under the other prevailing wind directions. Second, the city-scale cooperation on anthropogenic 351 emission reductions within the PRD should be enhanced under the westerly, southwesterly, 352 easterly, southeasterly, or southerly wind; whereas the emission control on a larger spatial scale 353 rather than only the PRD will be more effective under the northerly or northeasterly wind. 354 Finally, the prior control of local precursor emissions in conjunction with flexible reinforcing 355 upwind regional emission control based on prevailing wind direction is recommended to reduce 356 the O₃ levels and improve the O₃ attainment rate in GZ. 357

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473 **Figures and Tables**



475 Fig. 1. The flow scheme for response surface model based emission source contribution and

476 meteorological pattern analysis in ozone polluted days.

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479 Fig. 2. Three simulation domains and the regions in the d03 domain. GZ - Guangzhou; FS -

480 Foshan; ZS - Zhongshan; JM - Jiangmen; DG - Dongguan; SZ - Shenzhen; HZ – Huizhou;

481 OTH – The other regions in the d03 domain.



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Fig. 3. The 2-D isopleths of O_3 response in GZ to AVOC and NO_x emission changes from all source regions in the d03 domain and their matching PR value. The X and Y-axes show the emission ratio of AVOC and NO_x for the entire d03 domain; Different O_3 concentrations are presented in different colors.



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Fig. 4. Anthropogenic contribution within the d03 domain and the background contribution to
 O₃ concentrations in GZ. Base concentration: O₃ concentration under base scenario;
 Background concentration: O₃ concentration under background contribution; Different colors
 distinguish different prevailing wind directions.



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Fig. 5. The anthropogenic NO_x and VOC emission contributions from each source region in the d03 domain to O_3 of GZ under various prevailing wind directions.

Table 1. Synoptic patterns and their corresponding meteorological characteristics in O₃
 polluted days of GZ 2019.

Type (Season of occurrence)	Total days	The average value of MDA8 O ₃ concentration (Range) ^a	Mean temperature (Range) ^b	Mean relative humidity (Range) ^c	Mean wind speed (Range) ^d	Prevailing wind direction
Cyclone (Summer, Autumn)	26	200 (165-248)	31 (20-38)	52 (24-86)	1.34 (0.4-5.3)	West/ Southwest /South/ Southeast /North/ Northeast
Anticyclone (Spring, Summer, Autumn)	14	187 (163-228)	27 (17-34)	48 (25-76)	1.32 (0.5-3.5)	South/ Southeast/ North/East
Trough (Autumn)	9	190 (163-231)	29 (18-36)	54 (30-80)	1.25 (0.1-4.6)	Southeast/ North/ East
High pressure approaching to sea (Spring, Autumn)	5	192 (165-233)	26 (20-34)	51 (22-72)	1.08 (0.3-2.4)	South/ Southeast/ East

497 Unit: ^a µg/m³; ^b °C; ^c %; ^d m/s.